

See discussions, stats, and author profiles for this publication at: <https://www.researchgate.net/publication/335391344>

Succession of bacterial communities on carrion is independent of vertebrate scavengers

Preprint · August 2019

DOI: 10.1101/744748

CITATIONS

0

READS

115

5 authors, including:



Ethan Frehner

University of Utah

6 PUBLICATIONS 44 CITATIONS

SEE PROFILE



Evan Buechley

Smithsonian Institution

32 PUBLICATIONS 256 CITATIONS

SEE PROFILE



Cagan H Sekercioglu

University of Utah

272 PUBLICATIONS 9,149 CITATIONS

SEE PROFILE

Some of the authors of this publication are also working on these related projects:



Species Distribution and Biogeography of Convergent Nectarivores [View project](#)



Movement ecology of the Egyptian Vulture (*Neophron percnopterus*) [View project](#)

1

2 **Succession of bacterial communities on carrion is independent of vertebrate**
3 **scavengers**

4 Cody R. Dangerfield, Ethan Frehner, Evan Buechley, Çağan H. Şekerciöğlü, William J.
5 Brazelton

6 School of Biological Sciences, University of Utah, Salt Lake City, UT

7

8 Corresponding Author:

9 William J. Brazelton

10 School of Biological Sciences

11 257 South 1400 East Rm 201

12 Salt Lake City, UT 84112-0840 USA

13 william.brazelton@utah.edu

14

15 **Abstract**

16 The decomposition of carrion is carried out by a suite of macro- and micro-organisms
17 who interact with each other in a variety of ecological contexts. The ultimate result of carrion
18 decomposition is the recycling of carbon and nutrients from the carrion back into the ecosystem.
19 Exploring these ecological interactions among animals and microbes is a critical aspect of
20 understanding the nutrient cycling of an ecosystem. Here we investigate the potential impacts
21 that vertebrate scavenging may have on the microbial community of carrion. In this study, we
22 placed seven juvenile domestic cow carcasses in the Grassy Mountain region of Utah, USA and
23 collected tissue samples at periodic intervals. Using high-depth environmental sequencing of the
24 16S rRNA gene and camera trap data, we documented the microbial community shifts associated
25 with decomposition and with vertebrate scavenger visitation. The remarkable scarcity of animals
26 at our study site enabled us to examine natural carrion decomposition in the near absence of
27 animal scavengers. Our results indicate that the microbial communities of carcasses that
28 experienced large amounts of scavenging activity were not significantly different than those
29 carcasses that observed very little scavenging activity. Rather, the microbial community shifts
30 reflected changes in the stage of decomposition similar to other studies documenting the
31 successional changes of carrion microbial communities. Our study suggests that microbial
32 community succession on carrion follows consistent patterns that are largely unaffected by
33 scavenging.

34

35 **Introduction**

36 Carrion, or dead animal tissue, provides a nutrient-rich resource for a wide array of
37 organisms. At the smallest scale, both geographically and in organisms affected, carrion
38 contributes nutrients to soils via nutrient leaching, thereby affecting microbial communities in
39 soil near the carcass (Howard, Duos, and Watson-Horzelski 2010; Parkinson et al. 2009;
40 Parmenter and MacMahon 2009). The impacts of carrion can be seen more directly as a food
41 source to the many necrophagous arthropods and vertebrate scavengers (Jordan et al. 2015).
42 Moreover, larger-scale impacts of carrion have been well documented. For example, the massive
43 die-off of salmon and cicada lead to large increases in resources and nutrient availability that
44 affect a myriad of organisms including microbes, plants, fungi, and vertebrates (Hocking and
45 Reynolds 2011, 2012; Jordan et al. 2015; Tiegs et al. 2009, 2011; Yang 2004) The versatile
46 methods by which carrion can be produced and consumed gives it the potential to impact many
47 facets of an ecosystem, but the pathway by which it decomposes and enters the ecosystem's
48 nutrient cycle depends on the environmental conditions and the interactions that form between
49 the organisms that compete over its resources.

50 The decomposition of carrion occurs continuously (Schoenly and Reid 1987), but it
51 typically has a consistent progression that is categorized into five stages of decay based on
52 physical composition of the carcass: fresh, bloat, active decay, advanced decay, and putrid dry
53 remains (Payne 1965). A body enters the fresh stage at death when a depletion of internal oxygen
54 triggers autolysis of the cells. Concurrently, endogenous microbes then begin to metabolize the
55 body and produce volatile compounds. The carcass transitions to the bloat stage as the activity of
56 these microbes fill the body cavity with gases, causing the carcass to distend. Active decay
57 follows the bloat stage when the body cavity ruptures, releasing the gases and allowing

58 invertebrates to consume soft tissue within the body cavity. After the removal of most of the soft
59 tissue and decrease in invertebrate activity, the carcass transitions into the advanced decay stage.
60 The carrion enters the putrid dry remains stage when the carcass has desiccated and all that
61 remains are bones and small amounts of skin and hair (Goff 2009; Payne 1965).

62 The high nutrient content of carrion makes it a highly sought-after resource for many
63 organisms (Hanski 1987; Janzen 1977; Wilson and Wolkovich 2011). Due to the high level of
64 competition, the organisms that consume carrion have developed behaviors in order to
65 monopolize the nutrients of carcass for themselves. These complex interactions among microbes
66 and scavenging fauna, along with abiotic factors (e.g. precipitation and temperature), often
67 impact the duration and occurrence of the stages of decomposition (Carter, Yellowlees, and
68 Tibbett 2010, 2008; Comstock et al. 2015; Galloway, Jones, and Parks 1989; Payne 1965;
69 Rozen, Engelmoer, and Smiseth 2008; Shukla et al. 2017). After an animal's death, microbes
70 quickly colonize the carcass and begin to metabolize tissue while producing toxins in order to
71 hinder consumption from other organisms (Burkepile et al. 2006; Janzen 1977; Rozen et al.
72 2008). Vertebrate scavengers, especially vultures, seek to quickly locate and consume carcasses
73 before other scavengers consume them or before decomposition progresses (Buechley and
74 Sekercioglu 2016). Furthermore, some vultures, such as the Turkey Vulture (*Cathartes aura*),
75 have developed an unusually high tolerance to decomposer-produced toxins, such as botulism,
76 and the harsh conditions present in their hindgut reduce the likelihood of carrion microbes
77 surviving consumption and infecting the vulture itself (Beasley, Olson, and DeVault 2015;
78 DeVault, Rhodes Olin E., and Shivik 2003; Roggenbuck et al. 2014). Other behaviors such as
79 the burial of carcasses has developed in both vertebrate and invertebrate scavengers in order to
80 seclude the carrion from climatic conditions, microbes, and other scavengers to slow

81 decomposition and secure the resources for themselves (Frehner et al. 2017; Rozen et al. 2008;
82 Shukla et al. 2017). In addition to burying, Burying Beetles (*Nicrophorus spp.*) further suppress
83 competition with microbes by excreting antimicrobial exudates on the carcass. In doing so, these
84 beetles limit decomposition and can delay the carcasses from entering the bloat or active decay
85 stages (Shukla et al. 2017), which are mainly dictated by microbial and insect activity (Finley,
86 Benbow, and Javan 2015; Goff 2009; Payne 1965).

87 Researching these interactions is important to understand how carrion decomposition
88 impacts nutrient cycling and the importance that carrion has on ecosystems (Barton 2015; Barton
89 et al. 2013), and also to forensics, as the pattern of succession on carrion and cadavers can be
90 used to determine a postmortem time interval (PMI; Anderson 2015). Historically, a majority of
91 studies focused on forensic entomology to determine PMI (Byrd and Allen 2001; Michaud and
92 Moreau 2009; Payne 1965; Schoenly and Reid 1987). Recent studies have utilized DNA
93 sequencing technology to characterize the microbiome of carrion and investigate the potential
94 use of microbes as indicators for PMI (Guo et al. 2016; Hyde et al. 2013; Pechal et al. 2014,
95 2013). These studies have investigated the microbial communities associated with carrion
96 decomposition and how seasonal changes and macroinvertebrates impact those microbial
97 communities. Moreover, animal scavengers may also impact the microbial composition of the
98 carcass by introducing their own communities of microbes. Scavengers act as a vector of
99 dispersal for many microbes, so their presence or absence may have a significant impact on the
100 microbial community of carrion (Crippen, Benbow, and Pechal 2015).. However, to our
101 knowledge, no study has investigated the influence of scavenger activity on the microbial
102 community composition of carrion. In this study, we use environmental DNA sequencing and

103 vertebrate scavenging data to investigate decomposition dynamics and potential impacts that
104 vertebrate scavengers have on the microbiome of carrion in the Great Basin Desert of Utah.

105 **Methods**

106 **Study sites and field data collection**

107 In this study, we investigate the bacterial communities of bovine carcasses in the Grassy
108 Mountain region of Utah, USA (40.87°N, -113.03°W) from May to June, 2015. To do so, we
109 experimentally placed juvenile domestic cow (*Bos taurus*) carcasses (n=7) in the study site and
110 monitored their decomposition using camera traps to identify vertebrate scavenger activity and
111 by collecting tissue samples at regular intervals to identify progression of microbial
112 communities. The calves were obtained from one local Utah dairy and had died from natural
113 causes either during or shortly after birth. The carcasses were collected on the day of birth/death,
114 and were kept frozen until their placement in the field to minimize any decomposition
115 progression. They were placed at least 3 km apart and fixed to a concealed stake in the ground to
116 prevent scavengers from removing the complete carcass. The carcasses weighed between 18.6
117 and 26.9kg. Carcasses were placed on sites that included sparse Utah juniper (*Juniperus*
118 *osteosperma*), greasewood (*Sarcobatus vermiculatus*), and widely distributed cheatgrass
119 (*Bromus tectorum*). The soil in the study area is composed of loose to moderately compacted
120 limnological sediments, including gravels and clays. The study area is arid and largely
121 homogenous. Study area temperatures varied between 7-40°C, and there was no precipitation
122 during the experiment. We collected tissue samples from each of the carcasses during five
123 sampling periods (Day 1, Day 4, Day 12, Day 18, and Day 26) (Fig 1).

124 The carcasses were equipped with Bushnell Trophy Cam HD motion-activated cameras
125 to monitor vertebrate scavenging activity. The cameras were programmed to take 1 photo when
126 triggered, with a 10-s delay between subsequent photos to reduce saturation of photos from the
127 same animal visitation event. All photos collected over the course of the study were entered into
128 CameraBase Version 1.7 (Tobler 2007), a camera-trap photo management platform in Microsoft Access.
129 We analyzed each of these photos individually and identified any vertebrates present in the photos to
130 species. We identified arrival times after carcass placement and duration of presence at carcass for each
131 scavenger species.

132 Tissue samples were excised from hind-leg muscle tissue of each carcass. Soil samples
133 directly adjacent to where the carcass was placed were acquired during the first sampling period
134 for two of the seven sites. Soil samples for the remaining sites were taken 5 m from the carcass
135 during the second sampling period. DNA was extracted from all carcass and soil samples using
136 the PowerSoil DNA Isolation Kit (MO BIO Laboratories, Carlsbad, CA, USA) according to the
137 manufacturer's instructions and stored at -20°C .

138 **Bacterial 16S rRNA gene sequencing**

139 The samples were submitted to the Michigan State University genomics core facility for
140 bacterial 16S rRNA gene amplicon sequencing. The V4 region of the 16S rRNA gene (defined
141 by primers 515F/806R) was amplified with dual-indexed Illumina fusion primers as described by
142 Kozich et al. (2013). Amplicon concentrations were normalized and pooled using an Invitrogen
143 SequalPrep DNA Normalization Plate. After library QC and quantitation, the pool was loaded on
144 an Illumina MiSeq v2 flow cell and sequenced using a standard 500 cycle reagent kit. Base
145 calling was performed by Illumina Real Time Analysis (RTA) software v1.18.54. Output of RTA
146 was demultiplexed and converted to fastq files using Illumina Bcl2fastq v1.8.4. Paired-end

147 sequences were filtered and merged with USEARCH 8 (Edgar 2010), and additional quality
148 filtering was conducted with the mothur software platform (Schloss et al. 2009)) to remove any
149 sequences with ambiguous bases and more than 8 homopolymers. Chimeras were removed with
150 mothur's implementation of UCHIME (Edgar et al. 2011). The sequences were pre-clustered
151 with the mothur command pre.cluster (diffs =1), which reduced the number of unique sequences
152 from 1,136,609 to 784,953. This pre-clustering step removes rare sequences most likely created
153 by sequencing errors (Schloss and Westcott 2011).

154 **Bacterial diversity analyses**

155 The unique, pre-clustered sequences were considered to be the operational taxonomic
156 units (OTUs) for this study and formed the basis of all alpha and beta diversity analyses, as in
157 our previous study (Dangerfield, Nadkarni, and Brazelton 2017). Sequence reads were not
158 rarefied for alpha diversity and evenness calculations because there was no correlation between
159 diversity indices and sequencing depth for this study. Taxonomic classification of all sequences
160 was performed with mothur using the SILVA reference alignment (SSURefv123) and taxonomy
161 outline (Pruesse, Peplies, and Glöckner 2012). Taxonomic counts generated by mothur and
162 edgeR were visualized using the R package phyloseq 1.20.0 (McMurdie and Holmes 2013).

163 **Statistical analyses**

164 Alpha diversity and evenness were calculated with the Shannon, invsimpson, and
165 simpson even calculators provided in the mothur package (Schloss et al. 2009). Differences
166 between alpha diversities and evenness were tested for significance using the Dunnett-Tukey-
167 Kramer test, which accounts for multiple comparisons among samples with unequal sizes and
168 variances (Lau 2013). Beta diversity was measured using the Morista-Horn biodiversity index, as

169 implemented in mothur. This index was chosen because it reflects differences in the abundances
170 of shared OTUs without being skewed by unequal numbers of sequences among samples.
171 Differences between community compositions were tested for significance using AMOVA
172 (analysis of molecular variance) as implemented in mothur (Pruesse et al. 2012). Morisita-Horn
173 community dissimilarity among samples was visualized using a nonmetric multidimensional
174 scaling (NMDS) plot. This plot was generated using the ordinate and plot ordination commands
175 in phyloseq (McMurdie and Holmes 2013). The ggplot2 function `stat_ellipse` was added to draw
176 95% confidence level ellipses (assuming t-distribution) in the NMDS plot (Wickham 2016).
177 Environmental variables (temperature, scavenger counts, and scavenging duration) were fitted to
178 the community composition ordination with the `envfit` function in the `vegan` package (Oksanen et
179 al. 2019). Differences in the relative abundance of OTUs between stages was measured using the
180 R package `edgeR` (Robinson, McCarthy, and Smyth 2009) as recommended by McMurdie and
181 Holmes (2014). The differential abundance of an OTU (as measured in units of log₂ fold change
182 by `edgeR`) was considered to be statistically significant if it passed a false discovery rate
183 threshold of 0.05. Taxa were determined characteristic to that specific stage if they were found
184 differentially abundant in one stage compared to all other stages and soil. These taxa are referred
185 to as “characteristic taxa” for the purposes of this paper. To investigate potential environmental
186 contamination of carcass samples, OTUs with at least 20 sequence counts among all samples
187 were assigned to either fresh carcass or soil using the sink-source Bayesian approach of
188 `SourceTracker2` v2.0.1 (Knights et al. 2011) with rarefying to 66,001 sequences for sinks and
189 16,828 sequences for sources. Similar results were obtained without rarefying sequence counts.
190 The one carcass sample that was determined to be contaminated by soil via `SourceTracker2` was
191 excluded from alpha and beta diversity analyses.

192

193 **Data Availability**

194 All sequence data are publicly available at the NCBI Sequence Read Archive under BioProject
195 PRJNA525153.

196

197 **Results**

198 **Carcass decomposition**

199 Sampling periods were categorized into stages of decomposition based on physical
200 interpretation of the carcasses (Fig 1) as determined from camera trap photographs taken of each
201 carcass and as described by Payne (1965). Day 1 was determined to be “fresh”, as the carcasses
202 were kept frozen promptly after death. The carcasses entered the “bloat” stage in Day 4, as
203 evidenced by the body cavity becoming distended by gases emitted during microbial
204 decomposition. The later sampling periods (Day 12, 18, and 26) were all categorized as the
205 “active decay” stage, because the large decrease in carcass size and the presence of skin tissue on
206 the carcass.

207 Vertebrate scavengers that fed at the carcasses included American Badger (*Taxida taxus*),
208 Common Raven (*Corvus corax*), Coyote (*Canis latrans*), Kit Fox (*Vulpes macrotis*), Turkey
209 Vulture, and White-tailed antelope squirrel (*Ammospermophilus leucurus*). Turkey vultures were
210 the most frequent scavenger to feed at the carcasses, and the majority of vertebrate scavenging
211 occurred between Day 4 and Day 12 of the study (Table 2).

212

213 **Impact of scavenging on bacterial communities**

214 To identify the impact that scavenging had on bacterial community composition, we
215 compared Morisita-Horn dissimilarities between high and low scavenging sites directly after the
216 peak of scavenging. Sites 4 and 7 were considered to be the "high scavenging" sites as they
217 experienced 88% of the total scavenging duration observed during the study (Table 2).
218 Differences in bacterial community composition between low scavenging and high scavenging
219 sites were not significant, however. Additionally, fitting of scavenging parameters (individuals
220 per week and scavenging duration per week) to the NMDS ordination (Figure 2) yielded no
221 significant correlations.

222 To identify individual operational taxonomic units (OTUs) that scavengers may have
223 introduced to the carcasses, we contrasted the relative abundances of OTUs in high scavenging
224 sites to low scavenging sites directly after the major scavenging events using edgeR (Robinson et
225 al. 2009). This comparison yielded 39 OTUs that were differentially abundant in high
226 scavenging sites in comparison to low scavenging sites.

227 We also examined the relative abundance patterns of OTUs classified as genera reported
228 by previous studies to be associated with macroinvertebrates (Dharne et al. 2008; Gupta et al.
229 2014, 2011; Lee et al. 2014; Shukla et al. 2017; Tóth et al. 2008) All macroinvertebrate-
230 associated genera reach peak relative abundances during the later sampling periods except for
231 *Providencia* and *Myroides* (Fig. 3).

232 **Bacterial community changes over time**

233 Microbial community differences are visualized in Fig. 2, where each data point
234 represents the overall bacterial community composition of one sample and the distance between
235 points represents the dissimilarity between samples. There are two primary shifts in community

236 composition: one from Day 1 to Day 4 and a second from Day 4 to later sampling periods (Fig
237 2). The 95% confidence ellipses show consistent separation between these sampling periods. The
238 only exception to this pattern was a single sample from Day 4 that clusters with soil samples in
239 Figure 2. We examined the bacterial community composition of this outlier sample in more
240 detail with SourceTracker2 (Knights et al. 2011), which revealed that 83% of OTUs in the outlier
241 sample could be confidently assigned to soil. Therefore, we concluded that this sample had been
242 contaminated with soil during sampling and/or handling, and we excluded this sample from all
243 downstream analyses.

244 Figure 2 also shows that the bacterial community composition of all carcasses at later
245 sampling dates (Days 12, 18, 26) are highly similar and not significantly different from each
246 other. The patterns visualized in the NMDS ordination were tested with an AMOVA that
247 confirmed significant differences between the Day 1, Day 4, and later sampling periods (Table
248 1). These three clusters of bacterial community composition (Day 1, Day 4, and Days 12-18-26)
249 correspond to the three stages of decomposition identified by physical interpretation of the
250 carcasses (Day 1 = "fresh", Day 4 = "bloat", and Days 12-18-16 = "active decay").

251 Proteobacteria was the most common phylum in the fresh stage, accounting for 48% of
252 microbial community composition, decreasing to 11% for both the bloat and active decay stages
253 (Figure 4). Conversely, Firmicutes abundance increased as decomposition progressed. Firmicutes
254 increased from 31% in the fresh stage to 72% and 84% of the microbial community in bloat and
255 active decay, respectively. Moraxellaceae represented 30% of the total abundance of the fresh
256 stage (Fig 4), whereas Moraxellaceae only accounted for 2% and 0.8% of the total abundance in
257 bloat stage and active decay, respectively. By contrast, Clostridia was dominant in the bloat

258 stage, accounting for 70% of the total abundance, and accounting for only 3% of total abundance
259 in the fresh stage (Supplementary Krona files).

260

261 **Taxa characteristic to each stage of decomposition**

262 The taxonomic classifications (at order and family level) of all OTUs identified as
263 characteristic to each stage of decomposition (as defined by differential abundance analysis
264 described in Methods) are shown in Figure 4. Each bar in Fig. 4 represents the total community
265 composition of each stage, in which “Ubiquitous taxa” represents taxa that are equally abundant
266 across two or more groups and the remaining taxa in each bar represent OTUs that are
267 characteristic to that stage. There were 323 OTUs that were characteristic to the fresh stage, and
268 these taxa accounted for ~62% of the total abundance in that stage. These fresh stage taxa are
269 dominated by Moraxellaceae, Flavobacteriaceae, Pseudoalteromonadaceae, Planococcaceae,
270 Staphylococcaceae, and an unclassified Bacillales family. The bloat stage contained fewer
271 characteristic taxa (106 OTUs) than the fresh stage, and these taxa accounted for ~56% of the
272 total abundance in that stage. The characteristic bloat taxa mainly consisted of Clostridia OTUs.
273 In particular, five Clostridia OTUs accounted for ~49% of the bloat stage’s total abundance. The
274 active decay stage contained 230 characteristic OTUs that accounted for ~64 percent of the total
275 abundance. Active decay still contained OTUs from Clostridia, but these OTUs are different than
276 those Clostridia OTUs in the bloat stage and account for less of the total abundance. Active
277 decay stage contained other prominent characteristic taxa from Flavobacteriaceae,
278 Enterococcaceae, Xanthomonadales and an unclassified Bacillales family. All characteristic
279 OTUs, their proportional abundances, and taxonomical classifications are reported in the
280 Supplementary Materials.

281

282 **Alpha diversity**

283 Figure 5 reports the OTU Shannon diversity index of each stage of decomposition. Alpha
284 diversity is highest during the fresh stage and decreases during the bloat stage. Diversity
285 increases once the carcass enters the active decay stage. The shifts in Shannon diversity between
286 fresh and bloat stages and between bloat and active decay stages passed a Dunnett-Tukey-
287 Kramer significance test. Similar patterns were also evident with the Inverse Simpson and
288 Evenness (from Simpson) indices.

289

290 **Discussion**

291 **No correlation between vertebrate scavenging and bacterial community composition**

292 Carrion is a valuable resource for which many organisms compete, including vertebrate
293 scavengers, macroinvertebrates, and microbes. While there have been some studies documenting
294 the association of bacteria with macroinvertebrates on carrion (Pechal et al. 2013; Rozen et al.
295 2008; Shukla et al. 2017), this is the first study, to our knowledge, to investigate the impact of
296 vertebrate scavengers on the bacterial community of carrion. Our study site, in an arid region of
297 Utah was well-suited for this experiment because of its unusually sparse vertebrate scavenging
298 activity (Frehner et al. in prep). Studying the succession of bacterial communities in the absence
299 of vertebrate impacts would have been difficult or impossible in a typical environment with
300 vigorous vertebrate scavenging.

301 Surprisingly, our results indicate no perceptible impact of scavenging activity on the
302 bacterial community composition of carrion. While we did detect 39 OTUs that were more

303 abundant in sites with high scavenging rates, these OTUs were typically present in only a single
304 sample; none of them were consistently present in both high-scavenging sites. Overall, the
305 bacterial communities of carcasses that experienced vigorous vertebrate scavenging (e.g. 15
306 individuals and 3.2 *hours* of activity) were highly similar to the bacterial communities of
307 carcasses at the same stage of decomposition but with almost no vertebrate scavenging (e.g. 1
308 individual and 1 *minute* of activity). Additionally, we did not detect any bacterial taxa that were
309 consistently more abundant after peaks in scavenging activity. Because scavenging was so rare at
310 these study sites, it would have been necessary to investigate many more carcasses in order to
311 detect a consistent shift in the bacterial community in response to scavenger activity.
312 Furthermore, the impact of scavenging may be rapid and transient, so future studies should
313 consider sampling at much more frequent time points.

314 Although we saw no evidence of bacterial taxa consistently associated with vertebrate
315 scavenging, we did find taxa that have been previously associated with turkey vultures
316 (Roggenbuck et al. 2014). However, the abundances of these taxa were not higher in sites that
317 experienced more scavenging by turkey vultures. Furthermore, Roggenbuck et al. (2014)
318 speculated that these taxa are most likely derived not from the turkey vultures, but from the
319 carrion they consume.

320

321

322

323 **Bacterial taxa associated with macroinvertebrates**

324 Several genera of microbes that have been previously found to be associated with
325 scavenging macroinvertebrates were also present in our study (Dharne et al. 2008; Gupta et al.
326 2014, 2011; Lee et al. 2014; Shukla et al. 2017; Tóth et al. 2008). With the exception of
327 Providencia and Myriodes, most of these genera exhibited increased abundance when the
328 carcasses entered into active decay. This increase in macroinvertebrate-associated taxa is
329 consistent with the increases of macroinvertebrate activity that occur during the active decay
330 stage (Finley et al. 2015; Matuszewski et al. 2010; Payne 1965).

331

332 **Stages of decomposition have consistent bacterial communities**

333 During this study, the carcasses exhibited three stages of decomposition (fresh, bloat, and
334 active decay), and each stage was characterized by a unique community of bacteria. Although the
335 weather, geography, and vertebrate scavenging activity of our study was notably different than in
336 previous studies, our observed patterns of bacterial community succession are remarkably similar
337 to those previously reported for carrion (Hyde et al. 2013; Pascual et al. 2017; Pechal et al. 2014,
338 2013) Pechal et al. (2013) reported that Proteobacteria was dominant in the early stages of
339 decomposition and declined as decomposition progressed and that Firmicutes abundances
340 progressively increased in the later stages of decomposition. Similarly, 8 of the 14 families that
341 exhibited distinct temporal patterns during the decomposition process in Pascual et al. (2017)
342 exhibited similar patterns in our study (Fig 4).

343 Our results are also consistent with those of previous studies that have shown a rapid
344 disappearance of aerobic microbial communities during the bloat stage (Finley et al. 2015; Goff
345 2009; Hyde et al. 2013; Pascual et al. 2017). Abundance patterns of OTUs from Moraxellaceae,
346 all of which are known to be aerobic (Pascual et al. 2017), exemplify this pattern: they represent

347 30.3% of total abundance in the fresh stage, 2.4% in the bloat stage (when anaerobic Clostridia
348 and Enterobacteriaceae are dominant), and 0.78% in the active decay stage. Furthermore, the
349 dominance of anaerobic bacteria during the bloat stage is exemplified with Clostridia comprising
350 70% of the microbial community in the bloat stage (Supplementary material bloat krona file).
351 This dominance of bacteria community composition during the bloat stage may partially explain
352 why we observed the bloat stage having the lowest alpha diversity values, whereas Pascual et al.
353 (2017) showed highest alpha diversity during the bloat stage (described as the putrefaction
354 stage).

355 Most other studies observe more rapid progression through the stages of decomposition
356 (Pascual et al. 2017; Pechal et al. 2014, 2013). It is possible that carcasses during the last
357 sampling period of our study were in "advanced decay", but there was no discernible change in
358 tissue composition of the carcasses compared to the previous two sampling periods. The delay in
359 decomposition is most likely the result of climatic factors. Previous studies investigating carrion
360 microbial communities were conducted in areas with much higher humidity, whereas this study
361 experienced arid conditions and high temperatures during a period with no precipitation. Active
362 decay has been shown to be limited or hindered by hot and dry climates similar to the conditions
363 present in this study, resulting in partial mummification (Galloway et al. 1989). During this
364 process, carcasses develop a mummified shell over the skeleton as the skin desiccates while
365 macroinvertebrate activity continues underneath, in the body cavity. This partial mummification
366 may explain the prolonged intact composition of the carcass in the latter sampling periods (Fig 1)
367 associated with an increased abundance of macroinvertebrate-associated bacterial taxa (Fig 3).

368

369 **Conclusion**

370 In this study, we investigated microbial succession associated with decomposition of cow
371 carcasses that experienced notably little vertebrate scavenging activity. Our results demonstrate
372 that the few intense vertebrate scavenging events at some carcasses did not affect the bacterial
373 community of the carcass. Instead, the bacterial community composition of all carcasses
374 consistently reflected the stage of decomposition, regardless of vertebrate scavenging activity.
375 Our results are remarkably similar to those of other studies conducted in wetter, milder
376 conditions with greater vertebrate scavenging activity (Hyde et al. 2013; Pascual et al. 2017;
377 Pechal et al. 2013, 2014), suggesting that bacterial community succession on carrion follows
378 consistent patterns that are largely unaffected by many external factors.

379 Although our study did not detect any consistent effects from vertebrate scavengers on
380 whole bacterial communities of carrion, it remains possible that vertebrates could be important
381 vectors of bacteria between carcasses. Such effects are likely to be taxa-specific and may only be
382 detectable under controlled experimental conditions. Future studies could investigate this
383 possibility by sampling carcasses at greater temporal and spatial resolution, by conducting
384 tracing experiments with known vertebrate and microbial taxa, and by characterizing bacterial
385 communities of the skin, mouths, and digestive tracts of vertebrate scavengers.

386

387

388 **References**

389 Anderson, Gail S. 2015. "Human Decomposition and Forensics." Pp. 541–60 in *Carrion*

390 *Ecology, Evolution, and Their Applications*.

391 Barton, Philip S. 2015. "The Role of Carrion in Ecosystems." Pp. 273–86 in *Carrion Ecology*,

- 392 *Evolution, and Their Applications*.
- 393 Barton, Philip S., Saul A. Cunningham, David B. Lindenmayer, and Adrian D. Manning. 2013.
- 394 “The Role of Carrion in Maintaining Biodiversity and Ecological Processes in Terrestrial
- 395 Ecosystems.” *Oecologia* 171(4):761–72.
- 396 Beasley, James C., Zach H. Olson, and Travis L. DeVault. 2015. “Ecological Role of Vertebrate
- 397 Scavengers.” in *Carrion Ecology, Evolution, and Their Applications*.
- 398 Buechley, Evan and Cagan Sekercioglu. 2016. *The Avian Scavenger Crisis: Looming*
- 399 *Extinctions, Trophic Cascades, and Loss of Critical Ecosystem Functions*.
- 400 Burkepile, Deron et al. 2006. *Chemically Mediated Competition between Microbes and Animals:*
- 401 *Microbes as Consumers in Food Webs*.
- 402 Byrd, Jason H. and Jon C. Allen. 2001. “The Development of the Black Blow Fly, *Phormia*
- 403 *Regina* (Meigen).” *Forensic Science International* 120(1–2):79–88. Retrieved March 27,
- 404 2019
- 405 (<https://www.sciencedirect.com/science/article/pii/S0379073801004315?via%3Dihub>).
- 406 Carter, David O., David Yellowlees, and Mark Tibbett. 2010. “Moisture Can Be the Dominant
- 407 Environmental Parameter Governing Cadaver Decomposition in Soil.” *Forensic Science*
- 408 *International* 200(1):60–66. Retrieved
- 409 (<http://www.sciencedirect.com/science/article/pii/S0379073810001441>).
- 410 Carter, David O., David Yellowlees, and Mark Tibbett. 2008. “Temperature Affects Microbial
- 411 Decomposition of Cadavers (*Rattus Rattus*) in Contrasting Soils.” *Applied Soil Ecology*
- 412 40(1):129–37. Retrieved

- 413 (<http://www.sciencedirect.com/science/article/pii/S0929139308000590>).
- 414 Comstock, Jenna L., Jean Paul Desaulniers, H el ene N. LeBlanc, and Shari L. Forbes. 2015.
415 “New Decomposition Stages to Describe Scenarios Involving the Partial and Complete
416 Exclusion of Insects.” *Journal of the Canadian Society of Forensic Science* 48(1):1–19.
417 Retrieved (<http://dx.doi.org/10.1080/00085030.2014.929850>).
- 418 Crippen, Tawni L., M. Eric Benbow, and Jennifer L. Pechal. 2015. “Microbial Interactions
419 during Carrion Decomposition.” Pp. 31–63 in *Carrion Ecology, Evolution, and Their*
420 *Applications*.
- 421 Dangerfield, Cody R., Nalini M. Nadkarni, and William J. Brazelton. 2017. “Canopy Soil
422 Bacterial Communities Altered by Severing Host Tree Limbs.” *PeerJ* 5:e3773. Retrieved
423 (<https://peerj.com/articles/3773>).
- 424 DeVault, Travis L., Jr. Rhodes Olin E., and John A. Shivik. 2003. “Scavenging by Vertebrates:
425 Behavioral, Ecological, and Evolutionary Perspectives on an Important Energy Transfer
426 Pathway in Terrestrial Ecosystems.” *Oikos* 102(2):225–34. Retrieved
427 (<https://doi.org/10.1034/j.1600-0706.2003.12378.x>).
- 428 Dharne, M. S. et al. 2008. “Antibacterial Activities of Multi Drug Resistant Myroides
429 Odoratimimus Bacteria Isolated from Adult Flesh Flies (Diptera: Sarcophagidae) Are
430 Independent of Metallo Beta-Lactamase Gene.” *Brazilian Journal of Microbiology*
431 39(2):397–404.
- 432 Edgar, RC. 2010. “Search and Clustering Orders of Magnitude Faster than BLAST.”
433 *Bioinformatics* 26(19):2460–61.

- 434 Edgar, Robert C., Brian J. Haas, Jose C. Clemente, Christopher Quince, and Rob Knight. 2011.
435 “UCHIME Improves Sensitivity and Speed of Chimera Detection.” *Bioinformatics*
436 27(16):2194.
- 437 Finley, Sheree J., M. Eric Benbow, and Gulnaz T. Javan. 2015. “Microbial Communities
438 Associated with Human Decomposition and Their Potential Use as Postmortem Clocks.”
439 *International Journal of Legal Medicine* 129(3):623–32.
- 440 Frehner, Ethan H., Evan R. Buechley, Tara Christensen, and Çağan H. Şekercioğlu. 2017.
441 “Subterranean Caching of Domestic Cow (*Bostaurus*) Carcasses by American Badgers
442 (*Taxidea Taxus*) in the Great Basin Desert, Utah.” *Western North American Naturalist*
443 77(1):124–29.
- 444 Galloway, Alison, Allen Jones, and Bruce Parks. 1989. “Decay Rates of Human Remains in an
445 Arid Environment.” *Journal of Forensic Sciences* 34(3):607–16.
- 446 Goff, M. Lee. 2009. “Early Post-Mortem Changes and Stages of Decomposition in Exposed
447 Cadavers.” *Experimental and Applied Acarology* 49(1–2):21–36.
- 448 Guo, Juanjuan et al. 2016. “Potential Use of Bacterial Community Succession for Estimating
449 Post-Mortem Interval as Revealed by High-Throughput Sequencing.” *Scientific Reports*
450 6(April):1–11. Retrieved (<http://dx.doi.org/10.1038/srep24197>).
- 451 Gupta, A. K. et al. 2014. “Molecular Phylogenetic Profiling of Gut-Associated Bacteria in
452 Larvae and Adults of Flesh Flies.” *Medical and Veterinary Entomology* 28(4):345–54.
- 453 Gupta, Arvind Kumar et al. 2011. “*Ignatzschineria Indica* Sp. Nov. and *Ignatzschineria*
454 *Ureiclastica* Sp. Nov., Isolated from Adult Flesh Flies (Diptera: Sarcophagidae).”

- 455 *International Journal of Systematic and Evolutionary Microbiology* 61(6):1360–69.
- 456 Hanski, Ilkka. 1987. “Carrion Fly Community Dynamics: Patchiness, Seasonality and
457 Coexistence.” *Ecological Entomology* 12(3):257–66. Retrieved
458 (<https://doi.org/10.1111/j.1365-2311.1987.tb01004.x>).
- 459 Hocking, Morgan D. and John Reynolds. 2012. “Nitrogen Uptake by Plants Subsidized by
460 Pacific Salmon Carcasses: A Hierarchical Experiment.” *Canadian Journal of Forest
461 Research* 42:908–17.
- 462 Hocking, Morgan D. and John D. Reynolds. 2011. “Impacts of Salmon on Riparian Plant
463 Diversity.” *Science* 331(6024):1609–12. Retrieved
464 (<http://science.sciencemag.org/content/331/6024/1609>).
- 465 Howard, Gary T., Bronwyn Duos, and Erin J. Watson-Horzelski. 2010. “Characterization of the
466 Soil Microbial Community Associated with the Decomposition of a Swine Carcass.”
467 *International Biodeterioration and Biodegradation* 64(4):300–304. Retrieved
468 (<http://dx.doi.org/10.1016/j.ibiod.2010.02.006>).
- 469 Hyde, Embriette R., Daniel P. Haarmann, Aaron M. Lynne, Sibyl R. Bucheli, and Joseph F.
470 Petrosino. 2013. “The Living Dead: Bacterial Community Structure of a Cadaver at the
471 Onset and End of the Bloat Stage of Decomposition.” *PLoS ONE* 8(10):e77733. Retrieved
472 (<http://dx.plos.org/10.1371/journal.pone.0077733>).
- 473 Janzen, Daniel H. 1977. “Why Fruits Rot, Seeds Mold, and Meat Spoils.” *The American
474 Naturalist* 111(980):691–713. Retrieved (<http://www.jstor.org/stable/2460325>).
- 475 Jordan, Heather R., Jeffery K. Tomberlin, Thomas K. Wood, and M. Eric Benbow. 2015.

- 476 “Interkingdom Ecological Interactions of Carrion Decomposition.” Pp. 433–59 in *Carrion*
477 *Ecology, Evolution, and Their Applications*.
- 478 Knights, Dan et al. 2011. “Bayesian Community-Wide Culture-Independent Microbial Source
479 Tracking.” *Nature Methods* 8(9):761–63.
- 480 Kozich, James J., Sarah L. Westcott, Nielson T. Baxter, Sarah K. Highlander, and Patrick D.
481 Schloss. 2013. “Development of a Dual-Index Sequencing Strategy and Curation Pipeline
482 for Analyzing Amplicon Sequence Data on the MiSeq Illumina Sequencing Platform.”
483 *Applied and Environmental Microbiology* 79(17):5112–20.
- 484 Lau, M. K. 2013. *DTK: Dunnett-Tukey-Kramer Pairwise Multiple Comparison Test Adjusted for*
485 *Unequal Variances and Unequal Sample Sizes*.
- 486 Lee, Jae Kook et al. 2014. “Wohlfahrtiimonas Larvae Sp. Nov., Isolated from the Larval Gut of
487 Hermetia Illucens (Diptera: Stratiomyidae).” *Antonie van Leeuwenhoek, International*
488 *Journal of General and Molecular Microbiology* 105(1):15–21.
- 489 Matuszewski, Szymon, Daria Bajerlein, Szymon Konwerski, and Krzysztof Szpila. 2010. “Insect
490 Succession and Carrion Decomposition in Selected Forests of Central Europe. Part 1:
491 Pattern and Rate of Decomposition.” *Forensic Science International* 194(1–3):85–93.
- 492 McMurdie, Paul J. and Susan Holmes. 2013. “Phyloseq: An R Package for Reproducible
493 Interactive Analysis and Graphics of Microbiome Census Data.” *PloS One* 8(4):e61217.
- 494 McMurdie, Paul J. and Susan Holmes. 2014. “Waste Not, Want Not: Why Rarefying
495 Microbiome Data Is Inadmissible.” *PLoS Computational Biology*.
- 496 Michaud, Jean-Philippe and Gaétan Moreau. 2009. “Predicting the Visitation of Carcasses by

- 497 Carrion-Related Insects under Different Rates of Degree-Day Accumulation.” *Forensic*
498 *Science International* 185(1):78–83. Retrieved
499 (<http://www.sciencedirect.com/science/article/pii/S0379073808005057>).
- 500 Oksanen, Jari et al. 2019. “Vegan: Community Ecology Package.” Retrieved ([https://cran.r-](https://cran.r-project.org/package=vegan)
501 [project.org/package=vegan](https://cran.r-project.org/package=vegan)).
- 502 Parkinson, Rachel A. et al. 2009. “Microbial Community Analysis of Human Decomposition on
503 Soil.” Pp. 379–94 in *Criminal and Environmental Soil Forensics*, edited by K. Ritz, L.
504 Dawson, and D. Miller. Dordrecht: Springer Netherlands. Retrieved
505 (https://doi.org/10.1007/978-1-4020-9204-6_24).
- 506 Parmenter, Robert R. and James A. MacMahon. 2009. “Carrion Decomposition and Nutrient
507 Cycling in a Semiarid Shrub—Steppe Ecosystem.” *Ecological Monographs* 79(4):637–61.
508 Retrieved (<http://www.jstor.org/stable/40385231>).
- 509 Pascual, Javier et al. 2017. “Function of Bacterial Community Dynamics in the Formation of
510 Cadaveric Semiochemicals during in Situ Carcass Decomposition.” *Environmental*
511 *Microbiology* 19(8):3310–22.
- 512 Payne, Jerry A. 1965. “A Summer Carrion Study of the Baby Pig *Sus Scrofa* Linnaeus.” *Ecology*
513 46(5):592–602. Retrieved (<http://doi.wiley.com/10.2307/1934999>).
- 514 Pechal, Jennifer L. et al. 2013. “Microbial Community Functional Change during Vertebrate
515 Carrion Decomposition.” *PLoS ONE* 8(11):1–12.
- 516 Pechal, Jennifer L. et al. 2014. “The Potential Use of Bacterial Community Succession in
517 Forensics as Described by High Throughput Metagenomic Sequencing.” *International*

- 518 *Journal of Legal Medicine* 128(1):193–205.
- 519 Pruesse, Elmar, Jörg Peplies, and Frank Oliver Glöckner. 2012. “SINA: Accurate High-
520 Throughput Multiple Sequence Alignment of Ribosomal RNA Genes.” *Bioinformatics*
521 (*Oxford, England*) 28(14):1823–29.
- 522 Robinson, Mark D., Davis J. McCarthy, and Gordon K. Smyth. 2009. “EdgeR: A Bioconductor
523 Package for Differential Expression Analysis of Digital Gene Expression Data.”
524 *Bioinformatics* 26(1):139–40.
- 525 Roggenbuck, Michael et al. 2014. “The Microbiome of New World Vultures.” *Nature*
526 *Communications* 5:5498. Retrieved (<https://doi.org/10.1038/ncomms6498>).
- 527 Rozen, D. E., D. J. P. Engelmoer, and P. T. Smiseth. 2008. “Antimicrobial Strategies in Burying
528 Beetles Breeding on Carrion.” *Proceedings of the National Academy of Sciences*
529 105(46):17890–95.
- 530 Schloss, Patrick D. et al. 2009. “Introducing Mothur: Open Source, Platform-Independent,
531 Community-Supported Software for Describing and Comparing Microbial Communities.”
532 *Appl. Environ. Microbiol.* 75:7537–7541.
- 533 Schloss, Patrick D. and Sarah L. Westcott. 2011. “Assessing and Improving Methods Used in
534 Operational Taxonomic Unit-Based Approaches for 16S rRNA Gene Sequence Analysis.”
535 *Applied and Environmental Microbiology* 77:3219–26.
- 536 Schoenly, K. G. and W. Reid. 1987. “Dynamics of Heterotrophic Succession in Carrion
537 Arthropod Assemblages: Discrete Series or a Continuum of Change?” *Oecologia* 73:192–
538 202.

- 539 Shukla, Shantanu P., Heiko Vogel, David G. Heckel, Andreas Vilcinskas, and Martin
540 Kaltenpoth. 2017. “Burying Beetles Regulate the Microbiome of Carcasses and Use It to
541 Transmit a Core Microbiota to Their Offspring.” *Molecular Ecology* (March).
- 542 Tiegs, Scott D. et al. 2011. “Ecological Effects of Live Salmon Exceed Those of Carcasses
543 During an Annual Spawning Migration.” *Ecosystems* 14(4):598–614. Retrieved
544 (<https://doi.org/10.1007/s10021-011-9431-0>).
- 545 Tiegs, Scott D. et al. 2009. “Separating Physical Disturbance and Nutrient Enrichment Caused
546 by Pacific Salmon in Stream Ecosystems.” *Freshwater Biology* 54(9):1864–75. Retrieved
547 (<https://onlinelibrary.wiley.com/doi/abs/10.1111/j.1365-2427.2009.02232.x>).
- 548 Tobler, Mathias. 2007. “Camera Base.”
- 549 Tóth, Erika M. et al. 2008. “Wohlfahrtiimonas Chitinoclastica Gen. Nov., Sp. Nov., a New
550 Gammaproteobacterium Isolated from Wohlfahrtia Magnifica (Diptera: Sarcophagidae).”
551 *International Journal of Systematic and Evolutionary Microbiology* 58(4):976–81.
- 552 Wickham, Hadley. 2016. *Ggplot2: Elegant Graphics for Data Analysis*. Springer-Verlag New
553 York. Retrieved (<http://ggplot2.org>).
- 554 Wilson, Erin E. and Elizabeth M. Wolkovich. 2011. “Scavenging: How Carnivores and Carrion
555 Structure Communities.” *Trends in Ecology and Evolution* 26(3):129–35. Retrieved
556 (<http://dx.doi.org/10.1016/j.tree.2010.12.011>).
- 557 Yang, Louie H. 2004. “Periodical Cicadas as Resource Pulses in North American Forests.”
558 *Science* 306(5701):1565–67. Retrieved
559 (<http://science.sciencemag.org/content/306/5701/1565>).

560

561

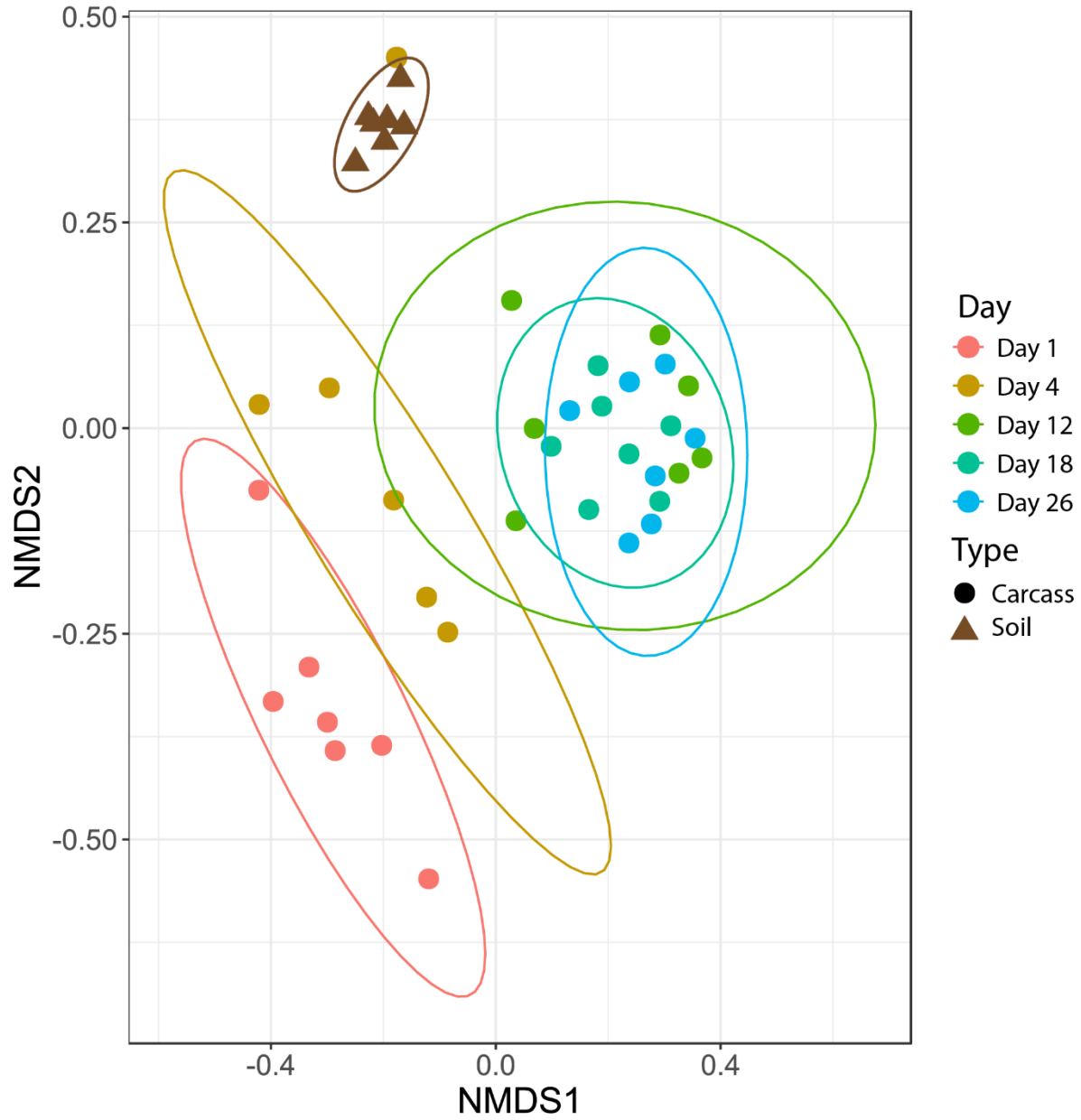
562 **Figures/Tables**



563

564 *Figure 1: Photographs of the decomposition progression at one of the seven sites. The photos from this site were selected for*
565 *clarity purposes, and all other sites had similar carcass composition.*

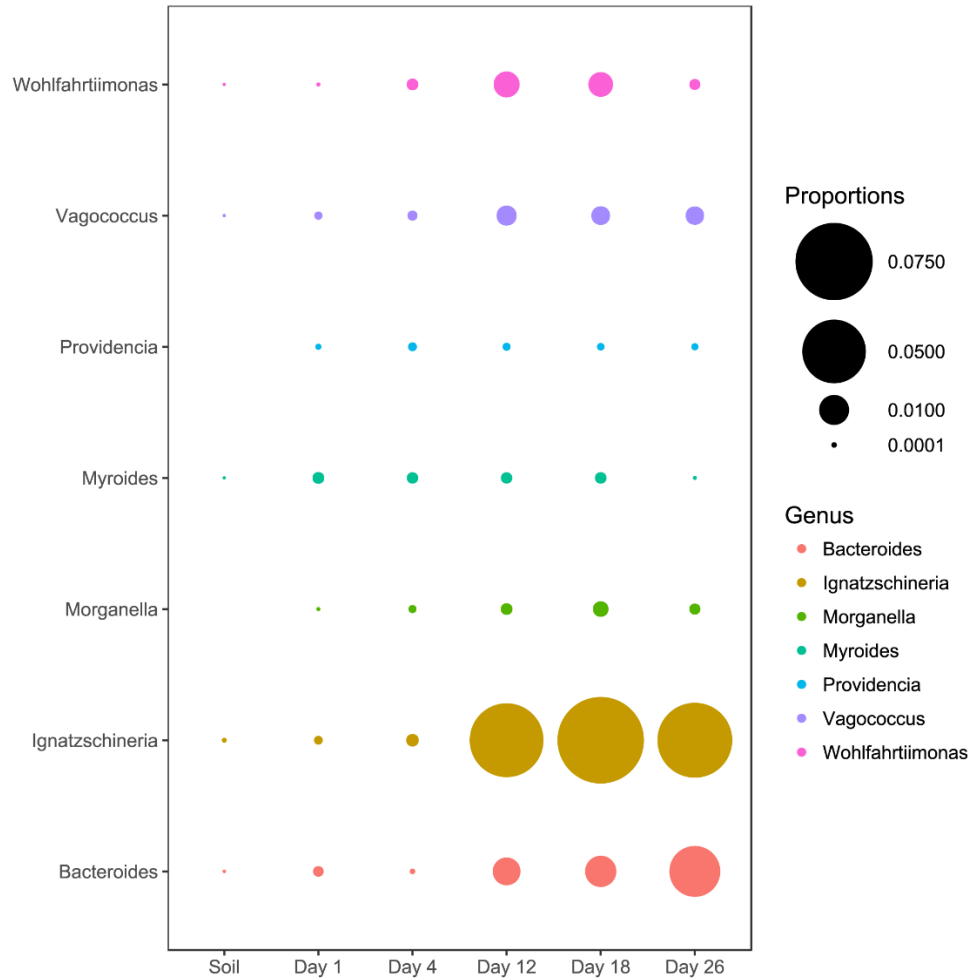
566



567

568 *Figure 2: Nonmetric multidimensional scaling (NMDS) plot showing bacteria community shifts associated with the stages of*

569 *decomposition. The ellipses indicate where 95% of samples within a sample period are expected to occur on the plot.*



570

571 *Figure 3: Abundance of genera associated with macroinvertebrates. All genera except Myroides and Providencia exhibited*
572 *abundance increases in the latter sampling periods.*

573

574

575

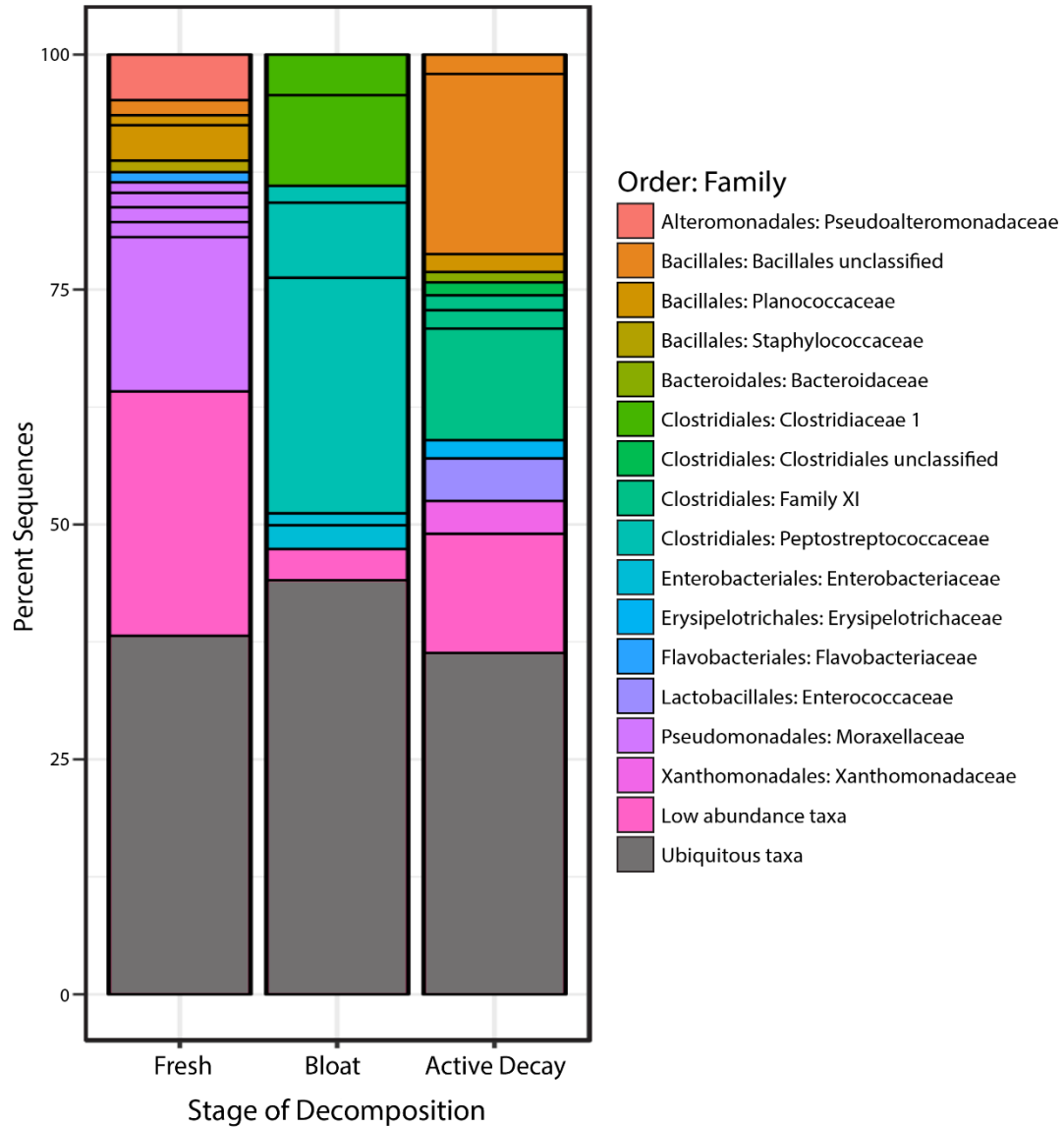
576

577

Description	Comparison	F_s	P-value
Overall	All 6 sample groups	9.30736	<0.001*
	Day 1 vs. Day 4	5.88847	0.001*
	Day 4 vs. Day 12	7.56111	<0.001*
	Day 12 vs. Day 18	0.406973	0.738
	Day 18 vs. Day 26	-0.22461	0.978
	Soil vs. Day 1	11.173	0.001*
	Soil vs. Day 4	7.47528	<0.001*
	Soil vs. Day 12	14.5593	<0.001*
	Soil vs. Day 18	16.8612	0.002*
	Soil vs. Day 26	16.8429	<0.001*

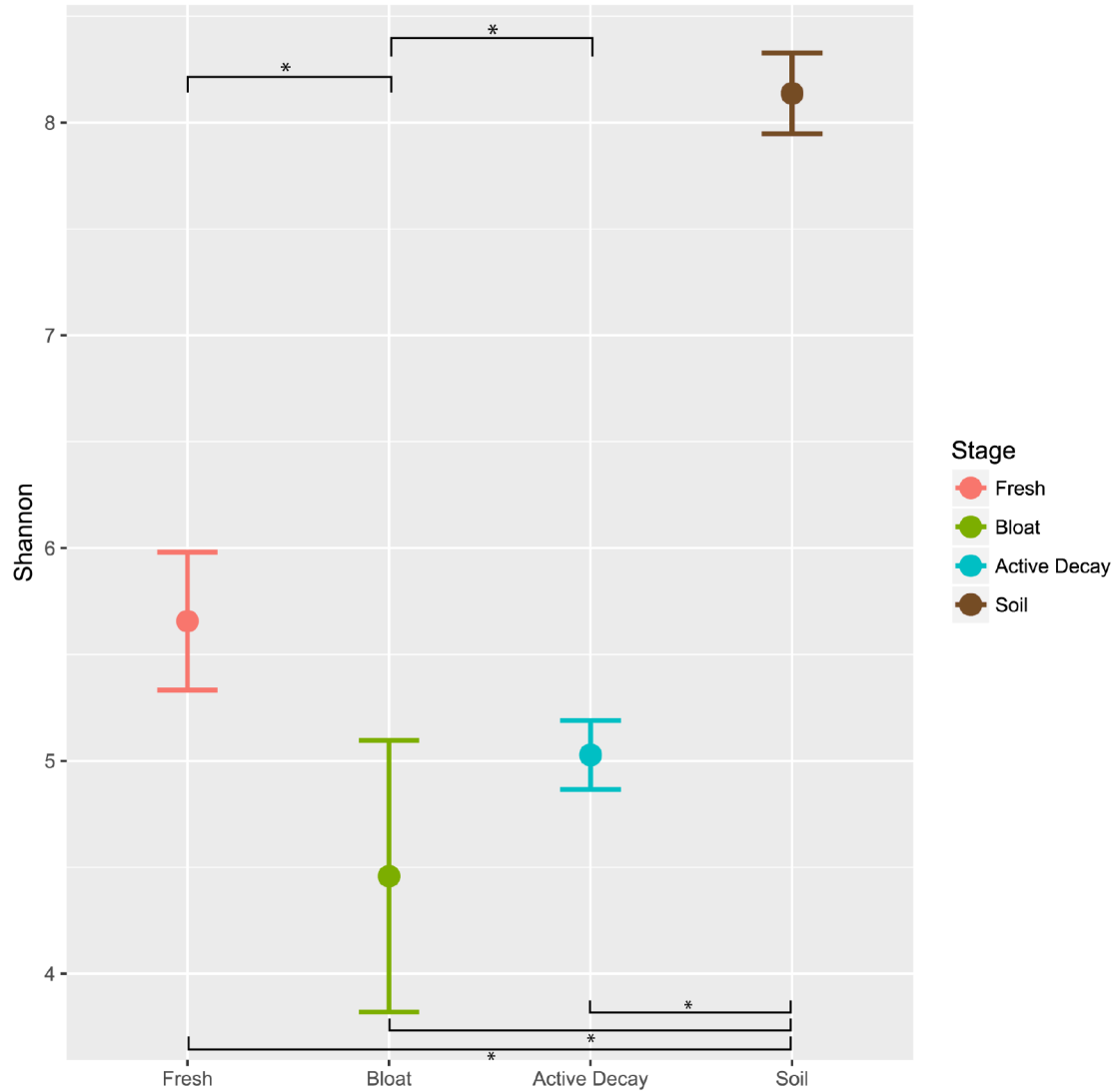
578 *Table 1: AMOVA analysis of significant differences in microbial community compositions. * = significant*

579



580

581 *Figure 4: Bacterial communities of each stage of decomposition. Taxa that are present throughout all stages or in soil is*
582 *represented by “Ubiquitous taxa”. The remain taxa are uniquely abundant within stage of decomposition. Uniquely abundant*
583 *taxa that had <1% abundance was grouped into “Low abundance taxa”.*



584

585 *Figure 5: Average Shannon alpha diversity and standard error between decomposition stages and soil. Significance of each*
586 *comparison was conducted by the Dunnett-Tukey-Kramer test. All comparisons were significantly different from one another*
587 *except fresh vs active decay. The red dot indicates an outlier (the sample that was contaminated by soil). * = significant ($p < 0.05$)*

588

589

590

591

592

593

Site	Species	Individuals	Scavenging Duration (Min)
Site 1	Coyote	1	3
	Turkey Vulture	1	9
	Total	2	12
Site 2	American badger	1	1
	Total	1	1
Site 3	Common Raven	1	4
	Coyote	1	1
	White-tailed Antelope Squirrel	6	22
	Total	8	27
Site 4	Common Raven	3	15
	Kit Fox	3	7
	Turkey Vulture	9	172
	Total	15	194
Site 5	Coyote	2	6
	White-tailed Antelope Squirrel	1	1
	Total	3	7
Site 6	White-tailed Antelope Squirrel	1	3
	Total	1	3
Site 7	Common Raven	16	101
	Coyote	1	1
	Turkey Vulture	4	85
	Total	21	187

594

595 *Table 2: Summary of scavenging activity per species per site.*

596

597

598